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Enhanced photocatalytic degradation performance of Bisphenol A over TiO₂-SiO₂ photocatalyst by improving specific surface area under simulated natural light irradiation

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ABSTRACT

The application of natural light in photocatalytic process is a potential energy source. In this study, TiO₂-SiO₂ nanoparticles with outstandingly enhanced photocatalytic activity have been fabricated by sol-gel method. The prepared photocatalysts with different molar ratio of TiO₂:SiO₂ as 100:0; 95:5; 85:15 and 75:25 were denoted as TiO₂, TS5, TS15 and TS25, respectively. Characterization of these photocatalysts was evaluated using transmission electron microscopy (TEM), X-ray diffraction (XRD), N₂ Adsorption and Desorption isotherms method (Brunauer-Emmet-Teller, BET). The specific surface area was improved in TiO₂-SiO₂ photocatalyst namely 193.9 m².g⁻¹ (at molar ratio of TiO_2 :SiO₂ of 95:5) compared to 33.6 m².g⁻¹ of TiO₂. The crystallite size was calculated around 5 nm from XRD data and uniform particle size distribution was observed in TEM image. The photocatalytic experiments were performed with bisphenol A (BPA) as model compound of organic pollutant. The effect of various operation parameters such as initial concentration, initial solution pH and photocatalyst dosage has been investigated. The kinetic studies of the photocatalytic degradation BPA over TS5 followed the pseudo-first order ($k=1.09\times10^{-2}$ min⁻¹) and degradation yield to be 82.4% BPA, at pH6.23, initial concentration to be 10 ppm and photocatalyst dosage to be 1.0 g.L⁻¹. The photocatalyst TS5 maintained activity after four cycles and remained 78%. The TiO₂-SiO₂ composite photocatalyst has shown to be a promising heterogeneous photocatalyst for organic degradation.

Introduction

In recent decades, environmental pollution has been receiving a great deal of public attention. Especially, the toxic organic compound in wastewater can affect directly to human health through water environment, thus some feasible solutions were developed to solve these problems such as photocatalysis [1-4], adsorption [5], adsorption – photocatalysis [6], photoelectrical oxidation [7], etc. Along with the development of hybrid semiconductor photocatalyst, titanium dioxide (TiO₂) as semiconductor has been studied as a base material to combine with other materials to improve its defects like small surface area, low adsorption of toxic organic compound and phase stability at high calcined temperature [8]. Therefore, the development for a http://doi.org/10.51316/jca.2020.069

material that can improve the TiO_2 surface area and phase stabilization needs to be studied.

The combination of SiO₂ also contributes significantly to increasing the specific surface area of the material, so that the ability to adsorb pollutants to the catalytic surface is enhanced. Many researchers have been reported their studies about TiO₂-SiO₂ hybrid material and achieved remarkable results. Vaithiyanathan Ramamoorthy and colleagues synthesized successfully TiO₂-SiO₂ composite photocatalyst, the composite photocatalyst 30% w/w TiO₂-SiO₂ have outstandingly enhanced specific surface area, namely 279 m².g⁻¹ as compare to 92 $m^2.g^{-1}$ of pure TiO₂. Accordingly, the percentage decolorizations of acid orange (AO20) were 67% and 89% for pure TiO_2 and 30% w/w TiO_2 -SiO₂, respectively. This result may be due to higher surface area [9]. In another research, Vinayak G. Parale and coworkers was synthesized hydrophobic TiO₂-SiO₂ composite aerogels via in-situ polymerization and solgel process, according that the molar ratio of TiO₂:SiO₂ to be 0.7:1.0. Furthermore, the specific surface areas were 62 m².g⁻¹ and 252 m².g⁻¹ for TiO₂ and TiO₂-SiO₂ composite, respectively. The degradation efficiency of methylene blue (MB) was obtained 57% and 87% after 90 min under UV irradiation for TiO₂ and TiO₂:SiO₂ composite [10], respectively. Moreover, the enhancing surface areas when combine SiO_2 and TiO_2 was also observed in many research [11-14]. It has been proved that the important role of SiO₂ in improvement specific surface of photocatalyst.

Bisphenol A ((CH₃)₂C(C₆H₄OH)₂, BPA), which is raw for plastic manufacture, especially material polycarbonate plastic and epoxy resins [15-17]. It can be discharged into the water and/or air environment during manufacturing. Besides, amount of residual BPA can release from plastic products to environment directly effecting to human health [15]. In recent decade, the application photocatalysis process in BPA degradation have been developed in many reports [18-21]. The photocatalytic material has many advantages, its structure can be doped with metal or nonmetal to enhance light adsorption intensity, BPA can be partially or completely degraded. Also, after BPA degradation, the photocatalyst can reuse after many uses. In the recent year, many researchers have reported their efforts to improve BPA removal efficiency. Particularly, Chao-Yin Kuo and colleagues synthesized C-doped TiO₂ (C-TiO₂) photocatalyst for degradation BPA under sunlight irradiation, as follow this research, the band gap energy of C-TiO₂ (molar ratio of C:Ti of 0.05:1) reduced significantly to 2.92 eV and the reaction rate

constant was estimated to be 2.08 h⁻¹. However, the specific surface of C-TiO₂ is quite low, just be 46.4 m².g⁻¹ event at low calcination temperature of 450 °C [22]. In another research, Shepherd Sambaza and coworkers studied BPA photocatalytic degradation on polyaniline (PANI) supported Ag@TiO₂ (Ag@TiO₂-PANI) nanocomposite. The Ag@TiO2-PANI composite was achieved band gap energy of 3.0 eV and the yield of BPA degradation up to 99.7% under visible light irradiation. This outstandingly achievement may be due to the presence of Ag nanoparticles on TiO₂ surface made recombination of electron and hole reduce. However, the specific surface of Ag@TiO₂-PANI was pretty low 52.5 m².g⁻¹ [20]. Finally, BPA degradation have been investigated in many researches, but the improvement in surface area, which is an important factor in photocatalytic process has not yet been focused.

In this research, we focus on enhancing the BET surface area of the photocatalytic process. The combination of SiO_2 with TiO_2 in TiO_2 -SiO_2 photocatalyst make the photocatalyst achieved large specific surface area a factor effect directly to adsorption, which is the first step of photocatalytic process. Furthermore, the presence of SiO₂ also improves phase stability. The photocatalytic activity experiments were tested in slurry reactor. Additionally, different operation parameters were investigated as initial BPA solution pH, photocatalyst dosage, initial BPA concentration. Besides, to increase practical applicability, the photocatalyst was undergone reuse experiment.

Experimental

Material

Reactants such as Tetraethyl orthosilicate (TEOS, \geq 99%, Merck, Germany), tetra-n-butyl orthotitanate (TNB, \geq 99%, Acros Organics), polyethylene glycol 20000 (PEG 20000, Sigma-Aldrich, Germany), acid nitric (HNO₃, 65%, Merck, Germany), absolute ethanol (C₂H₅OH, \geq 99.5%, Vietnam), acetyl acetone (AcAc, \geq 99%, Merck), Bis-phenol A (BPA, \geq 99%, Sigma-Aldrich) were used as received without any purified step.

Photocatalyst preparation

Photocatalyst materials TiO₂-SiO₂ were synthesized by sol-gel method. Firstly, solution included 4.00 g of TNB, 1.2 mL of AcAc, 5.0 mL of Ethanol and 0.5 mL of HNO₃ 65 wt.% was kept at 80 °C under magnetic stirring for 1h to form homogeneous solution A. At the same time, solution B including TEOS, 0.20 g of PEG, 5.0 mL of

Ethanol, 0.5 mL of HNO₃ 65 wt.% and 2.4 mL of deionized water was prepared under magnetic stirring. Amount of TEOS was designed so that the molar ratio of TiO₂:SiO₂ of 95:5; 85:5, 75:25 and denoted as TS5, TS15, TS25, respectively. Subsequently, solution A was added dropwise into solution B and kept stirring for 2h. Finally, TiO₂-SiO₂ solution was dried at 120 °C for 3h and then calcined at 500 °C for 2h. For comparison, TiO₂ powder was also synthesized without using TEOS.

Characterization

The phase compositions and crystal structures of the samples were determined by X-ray diffraction (XRD) using a D2 Phaser (Bruker, USA) with Cu-K α radiation ($\lambda = 0.154$ nm). The equipment operates at the step size of 0.02 ° at the rate of 0.2 °/s. The calculated crystallite size was determined by the Scherrer equation (Eq. 1) at (101) peak. In which β is the full width at half-maximum of the diffraction peak (FWHM), k=0.9 is a shape constant, D is the crystallite size and θ is the Bragg angle.

$$D = \frac{k\lambda}{\beta\cos\theta}$$
(1)

The N₂ adsorption-desorption isotherms were carried out at 77K by Quantachrome NOVA 1000E. The BET specific surface was determined from data over the relative pressure (P/P_o) ranging from 0.046 to 0.986. The TEM image was obtained by using FE-SEM JEOL 7401.

The zero-point charge (pH_{ZPC}) was determined as a previous study [23]. Firstly, 25 mL KCI 0.1N and diluted with distilled water to obtain 100 mL of solution with designed pH values (pH_i): 3, 4, 6, 7, 8, 9 by HCI 0.1N or NaOH 0.1N respectively. Secondly, the prepared solutions were added into the separated beaker, which contained 0.10 g of photocatalyst per each. The mixtures were stirred for 48h at speed of 180 rpm, then filtered with filter paper to remove the photocatalyst and recorded pH value (pH_f). The pH variation (Δ pH) was calculated as the following equation (Eq.2):

$$\Delta pH = pH_f - pH_i \quad (2)$$

Photocatalytic activity

The photocatalytic activities of photocatalysts were evaluated through the degradation of 200 mL BPA solution with initial concentration to be 10 mg.L⁻¹ under simulated natural light using a 25 W lamp (Natural light PT 2191-ExoTerra). The reactions were carried out in a slurry reactor, the suspension containing BPA solution

and photocatalysts was continuously stirred for 1h to achieve adsorption-desorption equilibrium. Following, the suspension has been irradiated to performing photocatalytic reactions. During this process, the suspension was stirred continuously and the temperature was maintained around 30 °C using a cooling water jacket. An aliquot suspension was extracted from the reactor at the designed time and separated using 0.45 µm springe filter for determining residual BPA concentration. For testing reusability, after each cycle, the photocatalyst was separated from the suspension by centrifugation, then dried at 120 °C and following calcined at 350 °C to evaporate residual BPA and other organic compounds before using for the next cycle. The residual BPA concentration was determined using Hitachi UV-Vis spectrometer equipment. The absorbance of BPA has been measured at a characteristic absorption wavelength of 225 nm. The BPA photocatalytic degradation efficiency was calculated as the following equation (Eq. 3):

$$H(\%) = \left(1 - \frac{A}{A_o}\right) \cdot 100\%$$
 (3)

where A_0 and A are the absorbance at 225 nm of characteristic wavelength of the BPA solution before and after irradiation, respectively.

To estimate the mass of photocatalyst needs for the photocatalytic experiment, the experiments were performed with different TS5 photocatalyst dosage which were 0.25 to 2.00 g.L⁻¹ of BPA solution with initial concentration of 10mg.L⁻¹ and natural pH to be 6.23. The effect of the initial concentration of BPA was investigated in the range from 10 to 40 mg.L⁻¹ and the photocatalyst dosage of 0.50 g.L⁻¹. To account for the effect of initial pH on the photocatalytic degradation of BPA on TS5 photocatalyst, the 10 mg.L⁻¹ of BPA solution was prepared and adjusted pH in the range from 4 to 8 and the photocatalyst dosage of 0.50 g.L⁻¹.

To investigate BPA photocatalytic degradation reaction kinetic, the BPA solution was determined concentration after every 30 min.

The Langmuir–Hinshelwood (L - H) kinetic model is expressed as Eq. (4) and is used to examine the kinetics of photocatalytic reactions [24]:

$$r=-\frac{dC}{dt}=\frac{k_{r}K_{a}C}{1+K_{a}C}\left(4\right)$$

where k_r (min⁻¹), K_a and C (mg/L) are the rate constant, Langmuir absorption constant and concentration of contaminant, respectively. At a very low concentration

of BPA (lower than 40 mg.L⁻¹), the solution can be seen as a diluted aqueous solution so KC <<1. The L-H equation can be simplified to pseudo-first order equation (Eq. 5) as follow:

$$\ln\left(\frac{C}{C_o}\right) = kt \quad (5)$$

where k is the reaction rate constant, C is the residual BPA concentration and the concentration, C_{\circ} is the concentration of BPA solution when the light was turned on.

Results and discussion

Figure 1 shows the XRD pattern of TiO₂-SiO₂ samples and pure TiO₂ were calcined at 500 °C for 2h. As shown, all identical peaks at 2 θ diffraction angles of 25.2° (101), 36.9° (004), 48.2° (200), 54.8° (211) and 63.0° (204) can be attributed to the anatase TiO₂ (Anatase – JCPDS-21-1272) and no characteristic peak of rutile was observed. Besides, no typical peak of SiO₂ was detected, this indicates that SiO₂ is present in an amorphous phase.



Figure 1: XRD pattern of pure TiO_2 and $\text{TiO}_2\text{-}\text{SiO}_2$ samples

The average crystallite size of the prepared particles is given in Table 1. Crystallite sizes of TiO_2-SiO_2 photocatalysts are smaller than that of pure TiO_2 and decrease with increasing SiO_2 concentration. This phenomenon is attributed to the formation of Si-O networks by excessive SiO_2 , which prevents the production of anatase crystallites because the TiO_2 in the anatase phase cannot contact with isomorphous octahedral site substitution to forming rutile phase [25].



Figure 2: (a) N₂ adsorption-desorption isotherms and (b) DA analysis of TiO₂, TS5, TS15 and TS25 photocatalysts

| Photocatalyst | Average crystallite size (nm) | SSA (m².g⁻¹) | Pore volume (mL.g ⁻¹) | Pore diameter (nm) |
|------------------|-------------------------------------|-----------------|--------------------------------------|-----------------------|
| TiO ₂ | 16 | 33.6 | 0.138 | 1.5 |
| TS5 | 6 | 193.9 | 0.158 | 1.2 |
| TS15 | 6 | 209.2 | 0.245 | 2.7 |
| TS25 | 5 | 247.8 | 0.336 | 2.0 |

Table 1: Crystallite size, BET surface and porous parameter of TiO₂ and TiO₂-SiO₂ samples

The microstructural properties of TiO_2-SiO_2 the photocatalyst with different molar ratios were investigated using N₂ adsorption-desorption isotherms at 77K and Dubinin-Astakhov (DA) method micropore analysis from the desorption profiles of TiO_2 and TiO_2 - SiO_2 photocatalysts. These samples indicated a type IV isotherm with a H3 hysteresis loop, characteristic of the combination of the microporous material. The BET surface areas were determined to be 33.6, 193.9, 209.2 and 247.8 m².g⁻¹ for TiO_2 , TS5, TS15 and TS25, respectively. The significant increase in specific surface area was observed in the presence of SiO_2 . The higher specific surface contributed multiple active sites and improved adsorption capacity, which are important in the photocatalytic process.

To clearly observe the crystal structure, the TEM images of TS5 and TS25 photocatalyst were taken and showed in Figure 3. The TEM images indicate that the TS5 and TS25 nanoparticles considered of spherical shape, and the average particle diameter was estimated about 10 - 15 nm and the uniform distribution of nanoparticles was observed. Although there may have been agglutination between the crystals, the TEM image still showed a smaller particle size than TiO₂ (16 nm), this can be attributed by the role of SiO₂ in photocatalyst. During the calcination process, the SiO₂ can be incorporated with TiO₂ and was forming gradually a layer coating on the TiO₂ surface, therefore, the crystallite size was reduced [8]. This phenomenon was consistent with the results obtained from XRD pattern.



Figure 3: TEM images of the prepared TiO₂-SiO₂ particles: (a) TS5 and (b) TS25

Effect of TiO₂:SiO₂ molar ratio

Figure 4 shows the comparison of photocatalytic activities of various TiO₂-SiO₂ photocatalysts. It can be seen from Figure 4, in the absence of the photocatalyst, BPA is stable under photolysis irradiation. During the photocatalytic process, almost photocatalyst can degrade BPA at any molar ratio, the TiO₂-SiO₂ catalyst indicated enhancing adsorption capacity rather than pure TiO₂ photocatalyst as a result of SiO₂ combination. After 240 min reaction, the experimental results show that the TS5 photocatalyst achieved the greatest photocatalytic activity, this result corresponding to previous research [26-29]. The decreased activity in TS25 photocatalyst may be due to the excess SiO₂ leaches out the photocatalyst, another reason can be attributed by SiO₂ overlay the activity site of TiO₂ at high content. This result shows an improvement in photocatalytic activity when combining a small amount of SiO₂.



Figure 4: Effect of various ratios of TiO₂:SiO₂ on photocatalytic degradation efficiency at C_{BPA}=10 mg.L⁻¹, photocatalyst dosage=1.0 g.L⁻¹, V_{BPA}=200 mL

Effect of photocatalyst dosage

The effect of photocatalyst dosage was carried out with TS5 photocatalyst in the range from 0.25 to 2.0 g.L⁻¹ at a BPA initial concentration of 10 mg.L⁻¹. The results were shown in Figure 5, according to that the photocatalyst dosage of 1.00 g.L⁻¹ shows the highest reaction rate constant and degradation efficiency. The phenomena can be explained when the photocatalyst dosage increase from 0.25 to 1.00 g.L⁻¹, increasing the amount of TS5 leads to a larger number of active sites that mean more than radicals were formed. Otherwise, increasing the photocatalyst dosage from 1.00 to 2.00 g.L⁻¹ reduces the degradation efficiency of BPA, this may be due to the increasing turbidity of the solution along with excess photocatalyst [30]. The increasing

turbidity makes photon flux cannot active the photocatalyst resulting in less than radicals.



Figure 5: Effect of TS5 dosage on BPA degradation under simulated natural light at C_{BPA} =10 mg.L⁻¹, V_{BPA} =200 mL, pH6.23

Effect of initial pH of BPA solution

The pH of the solution is an important parameter influence the adsorption behavior in the photocatalytic process. The pH value will determine the concentration of OH⁻ species, which is the precursor to the formation of hydroxyl radicals oxidizing agents in the photocatalyst process. The pH value also affects to the BPA form, when the pH lower than pKa (9.58) the BPA exists in its neutral form ((CH₃)₂C(C₆H₄OH)₂), on the contrary, the ionic BPA is formed ((CH₃)₂C(C₆H₄O⁻)₂) [31]. Besides, the surface charge properties determined by the zero-point charge enhances the interaction between photocatalyst and organic pollutant [32]. The experimental results determine the zero-point charge was shown in Figure 6, according to that the pH_{ZPC} was estimated to be 5.89.



Figure 6: TS5 zero-point charge estimation diagram

The experiment evaluated the effect of pH was carried out in the range from 4 to 8. The results were presented in Figure 7, the highest BPA degradation was documented at the original pH value of 6.23 as the comparison with the others pH values. This result can be explained by the acidic environment pH lower 5.89 make the TS5 surface charge is positive, while BPA exists in neutral form. The positively charged surface formation of TS5 is occupied with H⁺ ions, accordingly increasing the formation of hydroxyl radicals and increase BPA adsorption onto the photocatalyst surface may be due to electrostatic interaction between positive surface and phenolic ring in BPA structure [33]. As increasing pH values, when pH values higher than pH_{ZPC} the surface charge was inverted to negative, the adsorption BPA will be reduced cause both the photocatalyst surface and phenolic ring of BPA were negatively charged, and naturally, reduce BPA degradation efficiency. The reducing BPA degradation efficiency at a weak basic environment was similarly demonstrated in research of Blanco-Vega co-workers when study Ni-doped and TiO₂ photocatalyst [34].



Figure 7: (a) Effect of initial pH on BPA degradation under simulated natural light; (b) Plot of $-\ln(C/C_o)$ versus time for estimate reaction rate constant

Effect of initial concentration of BPA solution

The effect of the initial concentration of BPA on the degradation of BPA under the simulated solar light was later determined. Because BPA is a weak acid with two hydroxyl groups in structure, so the pH value was not significantly along changed increasing BPA concentration. The obtained results have been presented in Figure 8. The results indicated that the degradation efficiency and reaction rate constant reduce along with increasing BPA concentration. According to shown in Table 2, a significant change in the reaction was observed between two concentrations 20 and 30 mg.L⁻¹. The reason can be explained as two factors having a dependent effect on the photocatalytic process, the density of BPA was adsorption and the intensity of light absorbed on the TS5 surface. When the BPA concentration below 20 mg.L⁻¹, BPA molecules in solution will have good contact with photocatalyst and the photons can be easily absorbed by photocatalyst. However, when the BPA concentration above 30 mg.L⁻¹, the BPA molecules become excess lead to the photons were interrupted before reach with photocatalyst as consequently reduce the BPA degradation efficiency [35]. Similar results were demonstrated by Lin Yanyan and investigating colleagues when the effect of acetaminophen initial concentration to degradation performance on WO₃/TiO₂/SiO₂ composite under UV-Vis irradiation [29]. According to this research, the related development was recognized in a concentration range from 10 to 15 mg.L⁻¹. Hence the maximum suitable concentration for which the catalyst can well maintain the photocatalytic capacity is 20 mg.L⁻¹.

Figure 8: (a) Effect of BPA concentration on BPA degradation at pH6.23, photocatalyst dosage = 1.0 g.L⁻¹, V_{BPA} = 200 mL (b) Plot of -ln(C/C_o) versus time for estimate reaction rate constant

Figure 9: Recycling test of TS5 for photocatalytic degradation BPA at C_{BPA} =10 mg.L⁻¹, photocatalyst dosage=1.0 g.L⁻¹, pH6.23, V_{BPA}=200 mL

| Initial BPA concentration (mg.L ⁻¹) | Reaction rate constant (min ⁻¹) | R ² |
|---|---|----------------|
| 10 | 0.0109 | 0.990 |
| 20 | 0.0084 | 0.994 |
| 30 | 0.0048 | 0.996 |
| 40 | 0.0034 | 0.981 |

Table 2: Reaction rate constant for BPA degradation at different initial concentrations

Reusability test

To evaluate the reusability of TS5 photocatalyst, the experiments were performed on TS5 photocatalyst and the results of four times test were displayed in Figure 9. To prepare for next run, the weight loss would be compensated from an experiment that was run in parallel under the same conditions. Especially, after four times reuse, the degradation efficiency was remained 78%.

Conclusion

Through this research, TiO_2-SiO_2 photocatalyst was prepared, characterized and tested photocatalytic activity on degradation BPA under simulated natural light. The important roles of SiO_2 in phase stability were observed in the XRD pattern. As above synthesizing procedure, the $TiO_2:SiO_2$ photocatalyst with molar ratio of 95:5 achieved a large surface to be 193.9 and a small crystalline size of 6 nm. BPA was degraded 82.4% after 4h irradiation, and maintained 78% after four times reuse under initial concentration up to 20 mg.L⁻¹, natural pH of 6.23

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